

The use of natural and modified sepiolite for removal of arsenic in drinking water

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Abstract

The natural sepiolite clay mineral was obtained from General Directorate of Mineral Research and Exploration (MTA) of Turkey and modified with 2.5 M of NaOH. The natural and modified sepiolite samples were characterized by XRF, XRD, N₂ sorption, FTIR and SEM- EDS. The maximum As(III) and As(V) removal for natural sepiolite was achieved at an initial pH of 7 and 5.0, respectively; ca. 2.2 mg/g and 10.5 mg/g. The maximum As(III) and As(V) removal for modified sepiolite was achieved at an initial pH of 5.0; ca. 45.4 mg g⁻¹ and 38.4 mg g⁻¹, respectively. While the Langmuir isotherm fitted well with results obtained from the As(V) of SP and As(III) of Na-SP, the Freundlich isotherm model is more appropriate for the results obtained from As(V) of Na-SP.

Key words: Adsorption, arsenic, manganese, sepiolite,

1. Introduction

Sepiolite $(Si_{12})Mg_8O_{30}(OH)_6(OH_2)_4 \cdot 8H_2O$ (magnesium hydrosilicate) is a natural clay mineral and has various industrial applications such as adsorptive, rheologic, and catalytic applications [1, 2]. Sepiolite has a large capacity for adsorbing heavy metals, organic molecules, dyes and pesticides from wastewater due to its channels running parallel to the fiber axis [3]. The abundance of sepiolite mineral reserves as a raw material source in Turkey and its relatively low cost attract attention for its utilization in water treatment [4].

Studies [5-13] showed that the sepiolite has high adsorption capacity for the removal of heavy metal ions such as Cu^{2+} , Zn^{2+} , Cd^{2+} , Co^{2+} , Pb^{2+} , Sr^{2+} , and Cr^{6+} from water. However, there are a few studies for the removal of arsenic and manganese from drinking water by natural sepiolite [10, 14, 15].

The problem of arsenic contamination in natural waters is a global interest concerning the different field researchers. Arsenic is found as form of As (V) and As (III) in groundwater in many regions around the world. Arsenic poisoning in Japan, China, Mexico, Argentina, India, and USA has been reported [16-18]. A maximum contaminant level for arsenic has been recommended to reduce in drinking water as 10 μ g L⁻¹ by the World Health Organization (WHO). Therefore, research on improving the current methods or developing low-cost alternatives and new materials to remove arsenic from industrial effluents or drinking water is essential [19].

The aim of this study was to evaluate the capacity of natural and modified sepiolite to remove As(III) and As(V) from aqueous media via adsorption. The influence of the initial pH value, the initial metal concentration and the contact time were investigated. The chemical composition, morphology, phase composition, specific surface area and pore volume, and qualitative composition of the natural and modified sepiolite were determined.

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2. Materials and Method

2.1. Preparation of adsorbent

The natural sepiolite (SP) was obtained from General Directorate of Mineral Research and Exploration (MTA) of Turkey. The sepiolite was first ball milled to particle size in a range of 0.25-0.5 mm and dried in an oven at 120 °C overnight. The sepiolite was transformed to the Naform by a treatment with a 2.5 M NaOH solution at 80 °C. After washing and drying, the sample was calcined at 500 °C for 2 h and denoted as Na-SP.

2.2. Characterization of samples

The chemical composition of the natural and modified sepiolites was analyzed using Energy – dispersive X- Ray spectroscopy (EDS) (OXFORD INSTRUMENTS INCA X-Act/51-ADD0013) on a scanning electron microscope (SEM) (JEOL/ JSM-6610). The morphology of the natural and modified sepiolites was examined by a Scanning Electron Microscope (SEM) (JEOL/ JSM-6610). X-ray powder diffraction (XRD) patterns of the sepiolites were recorded on a Rigaku SmartLab X-ray diffractometer using non-monochromotographic Cu K α 1-radiation (40 kV, 40 mA, $\lambda = 1.5$). Scanning was in the range 5–65 °C of 2 θ . The specific surface area and micropore volume of the samples were measured using N₂ adsorption–desorption (AUTOSORB 1C) at – 196 °C. The surface area, total pore volume and micropore volume were determined by multipoint BET, t-plot and DR (Dubinin–Radushkevic), respectively. Infrared absorption measurements of the natural and modified sepiolites were carried out using a Fourier Transform Infrared (FTIR) spectrophotometer (Bruker Optics- Alpha). The FTIR spectra were obtained in the wavenumber range 650–4000 cm⁻¹ using single bounce ATR with diamond crystal.

2.3. Adsorption experiments

Batch adsorption experiments were carried out in glass flasks (0.1 L) using a magnetic shaker at 25 °C at a constant agitation of 200 rpm. In the kinetic studies, suspensions containing a range of $25-100 \text{ mg} \cdot \text{L}^{-1}$ arsenic were stirred for different periods of time at initial pH of 5. After the reaction, suspensions were centrifuged at 5000 rpm for 5 min in order to separate the solution and the solids. The initial and non-adsorbed concentrations of arsenic in supernatants were determined by Atomic Absorption Spectroscopy (AAS) (GBC A4382). All results were expressed as averaged values of duplicate tests.

The adsorption capacity $(q_{e_1} \text{ mg} \cdot \hat{g}^{-1})$ and removal percentage (%) of arsenic were determined using Eq.2.1 and 2.2.

$$qe = \frac{(Co - Ce)V}{m}$$
(1)

Removal efficiency (R, %) = $\frac{(Co-Ce)}{Co} \times 100$ (2)

where C_0 and Ce are the initial and final concentrations of arsenic (mg L⁻¹), *V* is the volume of solution (L) and *m* is the amount of adsorbent (g). Equilibrium data for natural and modified sepiolites were fitted to the Langmuir and Freundlich models as reported in reference [20].

3. Results and Discussion

3.1. Chemical composition of sepiolites

SEM-EDS analysis in Table 1 indicate that sepiolite is the major components along with traces of Al, K and Fe oxides. The treatment of the SP with NaOH leads to insignificant change of almost all cations except dealumination and desilication in SP. The removal of Mg^{2+} and formation of hydroxysodalite are significant.

| Sample | Mg | Al | Si | K | Na | Ca | Ti | Fe | Cl |
|--------|-------|------|-------|------|-------|------|------|------|------|
| SP | 20.08 | 9.45 | 66.97 | 0.46 | - | 0.97 | 0.20 | 1.68 | 0.05 |
| Na-SP | 14.76 | 4.24 | 27.12 | 0.55 | 46.86 | 1.84 | - | 4.42 | - |

Table 1. The composition of natural and modified sepiolites.

3.2. XRD results of samples

A comparison of XRD patterns of SP and Na-SP are shown in Fig. 1. The presence of peaks distinctive for sepiolite is obvious. After the treatment with NaOH, peaks at 21, 35° of 2 θ appeared because of the formation of hydroxysodalite as reported by Kang et al. [21]. Additionally, peaks at 7, 19 and 23° of 2 θ disappeared because of significant dealumination and desilication of the SP.

3.3.Surface area and pore size distribution of samples

The nitrogen adsorption and desorption isotherms of natural and modified sepiolites are shown in Fig. 3, and the surface characteristics calculated from isotherms are listed in Table 2. The shape of SP is consistent with Type I according to the I.U.P.A.C [25], which is characteristic of microporous materials with a plateau at high relative pressures. The BET surface areas of the SP and Na-SP samples demonstrated that treatment with NaOH caused the formation of mesopores in the solid particles, resulting in a lower surface area, revealing 187.8 m² g⁻¹ for SP and 7.09 m² g⁻¹ for Na-SP. The pore diameters change in the same direction, varying from 194 Å to 264 Å through significant desilication and dealumination.



Figure 1. XRD patterns of natural and modified sepiolites

3.4. SEM images of samples

SEM images of SP and Na-SP are shown in Fig. 2. After the NaOH treatment, the particles seemed to be melted and became very irregular, indicated that the severe desilication and dealumination may destroy the structure to some extent as seen in the XRD pattern [22]. NaOH treatment led to not only a decrease in particle size but also crystal deagglomeration as reported in references [23, 24].



Figure 2. SEM images of natural and modified sepiolites.

Table 2. Surface area and pore characteristics of natural and modified sepiolites

| Sample | $SA (m^2 g^{-1})^a$ | $V_{\rm T} ({\rm cm}^3 {\rm g}^{-1})^{\rm b}$ | $V_{\rm MP} (\rm cm^3 g^{-1})^c$ | $d(Å)^d$ |
|--------|---------------------|---|----------------------------------|----------|
| SP | 187.8 | 0.91 | 5.66x10 ⁻² | 194 |
| Na-NSP | 7.09 | 0.04 | 5.4×10^{-4} | 268 |

^a Multipoint BET method ; ^bVolume adsorbed at $p/p_0 = 0.99$.; ^cMicropore volume calculated by DR method ; ^d Average pore diameter



Figure 3. N₂ adsorption–desorption isotherm and differential pore size distribution of natural and modified sepiolites

3.5. FTIR of the samples

The FTIR spectra of the SP and Na-SP samples are shown in Fig. 4. FTIR bands of the SP and Na-SP samples can be observe in three regions [26, 27].

Bands in the 4000–3000 cm⁻¹ range correspond to the vibrations of the Mg–OH group (3690 cm^{-1}) , coordinated water (3568 cm^{-1}) and zeolitic water (at 3422 cm⁻¹); the intensity of these bands disappeared after the alkali treatment due to dealumination and removal of Mg cation based on EDS results.

The band at 1671 cm^{-1} is due to the vibration of zeolitic water, disappeared after the NaOH treatment.

Bands in the 1200–400 cm⁻¹ are related with silicate. The band centered at 1016 and 460 cm⁻¹ corresponds to Si–O–Si vibration that NaOH treatment not only shifts, which it lower

wavenumbers from 1016 cm⁻¹ to 876 cm⁻¹, but also decreases its intensity. Bands at 1215, and 980 cm⁻¹ correspond to Si–O bonds that they disappeared after NaOH treatment. Bands at 690 and 637 cm⁻¹ correspond to Mg–OH bond vibrations. After alkali treatment, the band intensity at 637 cm⁻¹ decreased and the band 690 cm⁻¹ disappeared because of removal of manganese as seen in EDS results.



Figure 4. FTIR spectra of the natural and modified sepiolites

3.6. Adsorption of arsenic by natural and modified sepiolites

3.6.1. Effect of initial pH

The initial pH value of the solution is an important controlling parameter in the arsenic adsorption process because it affects the chemistry of arsenic [30, 31] and surface properties of the adsorbent.

The dependence of the sorption capacity on the pH value is shown in Table 3. In the case of As (III) and As (V), maximal adsorption capacity of the SP is obtained at 1.5 of pH. With increasing initial pH value, the adsorption capacity of SP decreased because of an increase in negative charge of the adsorbent surface. On the other hand, maximal adsorption capacity of the SP for As (III) and As (V) were obtained at 5 of pH. This difference can be explained by the changed surface charge of natural sepiolite by NaOH treatment.

3.6.2. Sorption isotherms studies and modeling

Adsorption capacity of SP and Na-SP *with* adsorption time is shown in Fig. 6. The equilibrium time of As(V) on SP and Na-SP were determined as 100 min and 300 min respectively and the equilibrium time of As(III) on Na-SP was determined as 300 min. Slow adsorption rate of arsenic on Na-SP can be modification of the structure and surface with NaOH treatment.

The isotherm constants calculated from experimental data in Fig. 6 are shown in Table 3. For adsorption of As (V) on SP and As(III) on the Na-SP the correlation coefficients (\mathbb{R}^2) obtained from the Langmuir isotherm model were higher than those obtained from the Freundlich model and therefore, the Langmuir equation represents the adsorption processes better. Adsorption of As (V) on Na-SP showed opposite behavior. The maximum adsorption capacity of As(V) on the SP and Na-SP is 10.52 and 38.46, respectively that treatment of SP with NaOH increased almost four times its adsorption capacity for As(V). Because As(III) adsorption capacity of SP is low, the equilibrium time investigation for As(III) on SP was not studied. In contrast to SP, maximum adsorption capacity of Na-SP for As (III) is found as 45.4 mg g⁻¹.

Table 3. The influence of pH on the adsorption capacity (mg g^{-1}) of natural and modified sepiolites for As(III) and
As(V).

| Sample | SP | | Na-SP | | | |
|--------|---------|-------|---------|-------|--|--|
| pH | As(III) | As(V) | As(III) | As(V) | | |
| 1.5 | 8.92 | 26.84 | 5.68 | 5.07 | | |
| 3 | n.d | 1.77 | 2.99 | 0.78 | | |
| 5 | n.d | 13.48 | 46.70 | 35.96 | | |
| 7 | 2.15 | 2.62 | n.d | 0.11 | | |
| 9 | 2.05 | 1.45 | - | - | | |



Figure 6. The plot of arsenic ions removal rates against adsorption time (200 rpm, 0.02 g, 25 °C).

Table 4. Adsorption isotherm constants of As (III) and As (V).

| Sample | Langmuir isotherm | Freundlich isotherm | | | | | | |
|--------|----------------------|---------------------|----------------|----------------------|------|----------------|--|--|
| | $Q_{max}(mg g^{-1})$ | $B (L mg^{-1})$ | \mathbb{R}^2 | $k_F \ (L \ g^{-1})$ | 1/n | \mathbb{R}^2 | | |
| | As(V) | | | | | | | |
| SP | 10.52 | 0.0034 | 0.95 | 0.036 | 1.06 | 0.93 | | |
| Na-SP | 38.46 | 0.44 | 0.87 | 26.84 | 0.07 | 0.94 | | |
| | As(III) | | | | | | | |
| Na-SP | 45.4 | 0.13 | 0.89 | 18.17 | 0.18 | 0.63 | | |

Conclusions

NaOH treatment increases arsenic adsorption capacity for both arsenic species (III and V). While the Langmuir isotherm fitted well with results obtained from the As(V) of SP and As(III) of Na-SP, the Freundlich isotherm model is more appropriate for the results obtained from As(V) of Na-SP.

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References

- [1]Brauner K, Preisinger A. Struktur und Entstehung des Sepioliths. vol. 6; 1956.
- [2] Alvarez A. Sepiolite: properties and uses, vol. 37. Amsterdam: Elsevier; 1984.
- [3]Sotelo JL, Ovejero G, Rodriguez A, Alvarez S, Garcia J. Study of Natural Clay Adsorbent Sepiolite for the Removal of Caffeine from Aqueous Solutions: Batch and Fixed-Bed Column Operation. Water Air Soil Poll 2013; 224(3).
- [4]Balci S, Dincel Y. Ammonium ion adsorption with sepiolite: use of transient uptake method. Chem Eng Process 2002; 41(1):79-85.
- [5]Brigatti MF, Medici L, Poppi L. Sepiolite and industrial waste-water purification: Removal of Zn2+ and Pb2+ from aqueous solutions. Appl Clay Sci 1996; 11(1):43-54.
- [6]Brigatti MF, Lugli C, Poppi L. Kinetics of heavy-metal removal and recovery in sepiolite. Appl Clay Sci 2000; 16(1-2):45-57.
- [7]Kara M, Yuzer H, Sabah E, Celik MS. Adsorption of cobalt from aqueous solutions onto sepiolite. Water Res 2003; 37(1):224-232.
- [8] Alvarez-Ayuso E, Garcia-Sanchez A. Sepiolite as a feasible soil additive for the immobilization of cadmium and zinc. Sci Total Environ 2003; 305(1-3):1-12.
- [9]Vico LI, Acid-base behaviour and Cu2+ and Zn2+ complexation properties of the sepiolite/water interface. Chem Geol 2003; 198(3-4):213-222.

- [10]Kocaoba S. Adsorption of Cd(II), Cr(III) and Mn(II) on natural sepiolite. Desalination 2009; 244(1-3):24-30.
- [11] Bektas N, Agim BA, Kara S. Kinetic and equilibrium studies in removing lead ions from aqueous solutions by natural sepiolite. J Hazard Mater 2004; 112(1-2):115-122.
- [12]Lazarevic S, Jankovic-Castvan I, Jovanovic D, Milonjic S, Janackovic D, Petrovic R. Adsorption of Pb2+, Cd2+ and Sr2+ ions onto natural and acid-activated sepiolites. Appl Clay Sci 2007; 37(1-2):47-57.
- [13]Marjanovic V, Lazarevic S, Jankovic-Castvan I, Potkonjak B, Janackovic D, Petrovic R. Chromium (VI) removal from aqueous solutions using mercaptosilane functionalized sepiolites. Chem Eng J 2011; 166(1):198-206.
- [14]Ilic NI, Lazarevic SS, Rajakovic-Ognjanovic VN, Rajakovic LV, Janackovic DT, Petrovic RD. The sorption of inorganic arsenic on modified sepiolite: the effect of hydrated iron(III) oxide. J Serb Chem Soc 2014; 79(7):815-828.
- [15]Guerra DL, Batista AC, da Costa PCC, Viana RR, Airoldi C. Adsorption of arsenic ions on Brazilian sepiolite: Effect of contact time, pH, concentration, and calorimetric investigation. J Colloid Interf Sci 2010; 346(1):178-187.
- [16]Korte NE, Fernando Q. A Review of Arsenic(Iii) in Groundwater. Crit Rev Env Contr 1991; 21(1):1-39.
- [17]Bissen M, Frimmel FH. Arsenic a review. Part II: Oxidation of arsenic and its removal in water treatment. Acta Hydrob Hydrob 2003; 31(2):97-107.
- [18]Bissen M, Frimmel FH. Arsenic a review. Part 1: Occurrence, toxicity, speciation, mobility. Acta Hydroch Hydrob 2003; 31(1):9-18.
- [19]Mondal P, Majumder CB, Mohanty B. Laboratory based approaches for arsenic remediation from contaminated water: Recent developments. J Hazard Mater 2006; 137(1):464-479.
- [20]Ates A. Role of modification of natural zeolite in removal of manganese from aqueous solutions. Powder Technol 2014; 264:86-95.
- [21]Kang SJ, Egashira K, Yoshida A. Transformation of a low-grade Korean natural zeolite to high cation exchanger by hydrothermal reaction with or without fusion with sodium hydroxide. Appl Clay Sci 1998; 13(2):117-135.
- [22]Li YN, Liu SL, Zhang ZK, Me SJ, Zhu XX, Xu LY. Aromatization and isomerization of 1hexene over alkali-treated HZSM-5 zeolites: Improved reaction stability. Appl Catal a-Gen 2008; 338(1-2):100-113.
- [23]Melian-Cabrera I, Espinosa S, Groen JC, van de Linden B, Kapteijn F, Moulijn JA. Utilizing full-exchange capacity of zeolites by alkaline leaching: Preparation of Fe-ZSM5 and application in N2O decomposition. J Catal 2006; 238(2):250-259.
- [24]Melian-Cabrera I, Espinosa S, Mentruit C, Kapteijn F, Moulijn JA. Alkaline leaching for synthesis of improved Fe-ZSM5 catalysts. Catal Commun 2006; 7(2):100-103.
- [25]Janchen J, Morris RV, Bish DL, Janssen M, Hellwig U. The H2O and CO2 adsorption properties of phyllosilicate-poor palagonitic dust and smectites under martian environmental conditions. Icarus 2009; 200(2):463-467.
- [26]Cannings FR. An infrared study of hydroxyl groups on sepioliteJ Phys Chem, 1968; 72:1072-1074.

- [27] Serna C. AJL, Serratosa J.M. : Folding in sepiolite crystals. Clays Clay Miner 1975, 23: 452-457.
- [28] Kim YT, Jung KD, Park ED. Gas-phase dehydration of glycerol over ZSM-5 catalysts. Micropor Mesopor Mat 2010; 131(1-3):28-36.
- [29]Woolery GL, Kuehl GH, Timken HC, Chester AW, Vartuli JC. On the nature of framewor Bronsted and Lewis acid sites in ZSM-5. Zeolites 1997; 19(4):288-296.
- [30] Borah D, Satokawa S, Kato S, Kojima T. Sorption of As(V) from aqueous solution using acid modified carbon black. J Hazard Mater 2009; 162(2-3):1269-1277.
- [31]Ben Issa N, Rajakovic-Ognjanovic VN, Jovanovic BM, Rajakovic LV. Determination of inorganic arsenic species in natural waters-Benefits of separation and preconcentration on ion exchange and hybrid resins. Anal Chim Acta 2010; 673(2):185-193.